

## KINETIC AND EQUILIBRIUM STUDIES OF LEAD AND CADMIUM BIOSORPTION FROM AQUEOUS SOLUTIONS BY SARGASSUM *SPP.* BIOMASS

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### ABSTRACT

Contamination of the aqueous environment by heavy metals is a worldwide environmental problem. Biosorption of lead (II) and cadmium (II) from aqueous solutions by brown algae *Sargassum spp.* biomass was studied in a batch system. The heavy metals uptake was found to be rapid and reached to 88-96% of equilibrium capacity of biosorption in 15min. The pseudo second-order and saturation rate equations were found in the best fitness with the kinetic data ( $R^2 > 0.99$ ). The data obtained from experiments of single-component biosorption isotherm were analyzed using the Freundlich, Langmuir, Freundlich-Langmuir and Redlich-Peterson isotherm models. The Redlich-Peterson equation described the biosorption isotherm of  $Pb^{2+}$  and  $Cd^{2+}$  with high correlation coefficient ( $R^2 > 0.99$ ) and better than the other equations. The effect of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  on the biosorption of  $Pb^{2+}$  was not significant, but the metal ions affected the biosorption of  $Cd^{2+}$  considerably. According to the Langmuir model, the maximum uptake capacities ( $q_m$ ) of *Sargassum spp.* for  $Pb^{2+}$  and  $Cd^{2+}$  were obtained as 1.70 and 1.02mmol/g, respectively. Although the *Sargassum spp.* used in this study can be classified as an efficient biosorbent.

**Key words:** Biosorption, sargassum, lead (II); cadmium (II), kinetic, isotherm

### INTRODUCTION

Aqueous heavy metals pollution represents an important environmental problem due to their toxic effects and accumulation throughout the food chain. Among heavy metals, lead and cadmium have high priority for removal from aqueous environments (Kapoor *et al.*, 1999; Volesky, 2001; Rama *et al.*, 2002 and Anonymous, 2002). The conventional technologies for the removal of heavy metals from wastewater mainly include: chemical precipitation, ion exchange, adsorption, membrane processes and evaporation that require high capital investment and running costs (Anonymous, 2000; Gupta *et al.*, 2001 and Aksu, 2002). Therefore, there is an urgent need for development of innovative but low cost processes, where metal ions can be removed economically. The search for new treatment technologies has focused on

biosorption (Dönmez *et al.*, 1999; Figueira *et al.*, 2000 and Loukidou *et al.*, 2003). Biosorption is a term that describes the removal of heavy metals by the passive binding to nonliving microorganisms (bacteria, fungi and algae) and other biomass (such as peat, rice hull, fruit peel, leave and bark of tree etc.) from an aqueous solution (Davis *et al.*, 2003 and Ma and Tobin, 2003). A number of different metal binding mechanisms have been postulated to be active in biosorption such as ion exchange, complexation, coordination, chelation, physical adsorption and microprecipitation (Volesky, 2001). Biosorption has many advantages including low capital and operational costs, the selective removal of metals, biosorbent regeneration and metal recovery potentiality, rapid kinetics of adsorption and desorption and no sludge generation. Biosorption technology has been shown to be a feasible alternative for removing heavy metals from wastewater. This technology can utilize naturally

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abundant biomass such as seaweeds, and of these *Sargassum* has been identified for its high sorption capacity (Volesky, 2001; Davis *et al.*, 2003 and Diniz and Volesky, 2003). Biosorption of lead (II) and cadmium (II) from aqueous solutions using various biomasses has been studied. Cadmium (II) biosorption on *Aspergillus oryzae* reached equilibrium in 1h with 90% of biosorption taking place in the initial 10 min (Kiff and Little, 1986). Kinetic data of cadmium (II) biosorption by chitin presented high correlation with the pseudo second order rate equation (Benguella and Benaissa, 2002). Matheickal and Yu (1999) observed that the maximum uptake capacities of *Durvillaea potatorum* and *Ecklonia radiata* for  $Pb^{2+}$  were 1.6 and 1.3 mmol/g, respectively.

The objective of this research was to study the biosorption of lead (II) and cadmium (II) from aqueous environments by brown algae *Sargassum spp.* biomass. Kinetic and isotherm of lead (II) and cadmium (II) biosorption and the effects of pH and light metal ions ( $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$ ) on lead (II) and cadmium (II) biosorption were investigated.

## MATERIALS AND METHODS

### *Preparation of biosorbent*

The biosorbent used in experiments was brown algae *Sargassum spp.* biomass. The biomass was harvested from Oman Sea on the coast of Chabahar, Iran. The biomass was washed with tap water and deionized water to remove sand and other impurities. The biomass was sundried and then dried in an oven at 70 °C. Dried biomass was ground in a laboratory blender. After this, the biomass was sieved to select particle between 0.2-0.3 mm for use. The biomass was subsequently loaded with  $H^+$  in a solution of 0.1M HCl (biomass concentration of 50 g/L) for 30 min under slow stirring. Later the biomass was washed with deionized water to remove excess hydrogen ions. Finally the biosorbent again dried at 70 °C for 24 hr.

### *Chemicals*

Synthetic solutions were prepared using de-ionized water and salts of  $Pb(NO_3)_2$ ,  $Cd(NO_3)_2 \cdot 4H_2O$ , NaCl, KCl,  $MgCl_2 \cdot 6H_2O$  and  $CaCl_2 \cdot 2H_2O$  (Merck

supplied). Initial pH of solutions was adjusted with a pH meter (CAMLAB Ltd, Model CG842) to the desired values by using 0.1-1M HCl and 0.1-1M NaOH.

### *Biosorption experiments*

In all batch biosorption experiments, solution volume was 1L and the mixture of solution and biosorbent was agitated in 200 rpm. The experiments were conducted at room temperature ( $20 \pm 1$  °C). Initial pH of the solutions was adjusted to desired values. The reaction mixture pH was not regulated after the initiation of experiments and final pH was measured.

### *Kinetic experiments*

Kinetic experiments were done in three initial concentrations of  $Pb^{2+}$  and  $Cd^{2+}$  and fixed initial ratio of adsorbate to biosorbent inside a single component system. Initial metal concentrations were 0.5, 1 and 5 mM and initial ratio of adsorbate to biosorbent was 2 mmol/g, therefore 2.5, 0.5 and 0.25g of the biosorbent were added to experiment vessels with initial metal concentrations of 5, 1 and 0.5 mM, respectively. Initial pH of solutions was adjusted to 5 and pH of the solutions was monitored continuously. The experiments were continued for 5 hr and samples were drawn from the mixture at predetermined time intervals for analysis.

### *Equilibrium experiments*

Kinetic experiments presented that maximum time required to reach equilibrium was 2 h; therefore, the equilibrium time for equilibrium experiments was chosen 3 hr.

### *Biosorption isotherm*

Biosorption isotherm experiments were conducted in a single component system. The initial  $Pb^{2+}$  and  $Cd^{2+}$  concentrations were varied from 0.05 to 5 mM. Initial pH of the solutions was adjusted to 5 and then 500 mg of *Sargassum spp.* biomass was added to experiment vessels.

### *Effect of pH on biosorption*

The effect of pH on equilibrium capacities of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption was studied in a single component system. Initial heavy metal ions concentration was 1 mM and initial pH of solutions

was varied from 2 to 5.5. After pH adjustment, 500 mg of *Sargassum* spp. biomass was added to experiment vessels.

*Effect of light metal ions on biosorption*

The effect of Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup> and Ca<sup>2+</sup> on equilibrium uptake of Pb<sup>2+</sup> and Cd<sup>2+</sup> was studied in a binary system (one heavy metal and one light metal). Initial heavy metal ion concentration was 1mM and initial light metal ion concentration was varied from 0 to 6 mM. Initial pH of the solutions was adjusted to 5 and then 500 mg of *Sargassum* spp. biomass was added to experiment vessels.

*Metal analysis:* The biomass was removed by filtration through 0.45µm membrane filters (mixed cellulose ester) and filtrates were analyzed for residual heavy metal (Pb<sup>2+</sup> or Cd<sup>2+</sup>) concentration by a flame atomic absorption spectrophotometer (FAAS, Chem. Tech Analytical, Model ALPHA4).

*Kinetic modeling:* Kinetic of Pb<sup>2+</sup> and Cd<sup>2+</sup> biosorption was modeled by the pseudo first-order (Langergren), pseudo second-order, saturation (mixed-order) and second-order rate equations presented below as Eqs. (1)-(4), respectively:

$$\ln \frac{(q_e - q_t)}{q_e} = -k_1 t \quad (1)$$

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (2)$$

$$\frac{1}{t} \ln \frac{C_0}{C_t} = -\frac{k_0}{K} - \frac{1}{K} \left( \frac{C_0 - C_t}{t} \right) \quad (3)$$

$$\frac{1}{(q_e - q_t)} = \frac{1}{q_e} + kt \quad (4)$$

where  $q_e$  and  $q_t$  are the amounts of metal ion sorbed (mmol/g) at equilibrium and at any time, respectively;  $k_1$  is the pseudo first-order rate constant of adsorption (min<sup>-1</sup>);  $k_2$  is the pseudo second-order rate constant of adsorption (gm/mol/min);  $C_0$  and  $C_t$  are the concentrations of metal ion (mM) at t=0 and at any time, respectively;  $k_0$  (mM/min) and  $K$  (mM) are saturation rate

constants of adsorption and  $k$  is the second-order rate constant of adsorption (gm/mol/min) (Benguella and Benaissa, 2002; Metcalf & Eddy, 2003; Azizian, 2004).

*Isotherm modeling*

The isotherm of Pb<sup>2+</sup> and Cd<sup>2+</sup> biosorption was analyzed using the Freundlich, Langmuir, Freundlich-Langmuir and Redlich-Peterson models.

The empirical Freundlich model based on sorption onto a heterogeneous surface is given below by Eq. (5):

$$q_e = K_F C_e^{1/n} \quad (5)$$

where  $C_e$  is equilibrium concentration of metal ion (mM);  $K_F$  and  $n$  are indicators of biosorption capacity and biosorption intensity, respectively (Loukidou *et al.*, 2004; Selatnia *et al.*, 2004b).

The Langmuir equation is based on the assumption that maximum adsorption corresponds to a saturated monolayer of solute on the adsorbent surface, that energy of adsorption is constant and that there is no transmigration of adsorbate in the plane of the surface. The Langmuir equation is given by Eq. (6):

$$q_e = \frac{b q_m C_e}{1 + b C_e} \quad (6)$$

where  $q_m$  is the maximum capacity of biosorption (mmol/g) and  $b$  is a constant related to the affinity of the binding sites (mL/mol) (Yalçýnkaya, *et al.*, 2002; Sheng, *et al.*, 2004).

The three-parameter Freundlich-Langmuir model was developed to improve the fitness found by the Freundlich or Langmuir model. This model is given by Eq. (7):

$$q_e = \frac{b q_m C_e^{1/n}}{1 + b C_e^{1/n}} \quad (7)$$

where  $b$ ,  $q_m$  and  $n$  are the Freundlich-Langmuir parameters (Volesky, 2003).

The three-parameter Redlich-Peterson model is given below by Eq. (8):

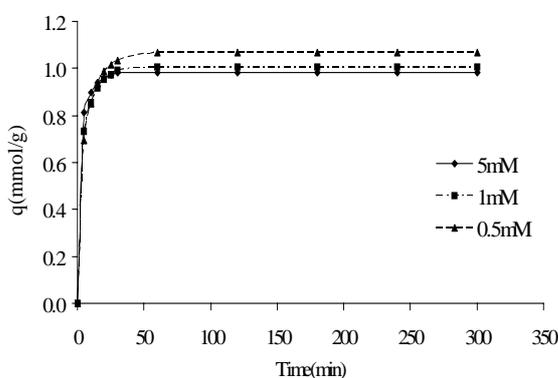
$$q_e = \frac{K_{RP} C_e}{1 + a_{RP} C_e^\beta} \quad (8)$$

where  $K_{RP}$  (L/g),  $a_{RP}$  (Lmmol<sup>-1</sup>) <sup>$\hat{a}$</sup>  and  $\hat{a}$  (dimensionless) are the Redlich-Peterson constants.  $\hat{a}$  lies between 0 and 1. For  $\hat{a} = 1$  the Redlich-Peterson model converts to the Langmuir model (Aksu, 2002 and Volesky, 2003).

## RESULTS

### Kinetic study

The kinetic profiles of Pb<sup>2+</sup> and Cd<sup>2+</sup> biosorption by *Sargassum spp.* were shown in Fig. 1(a)-(b).

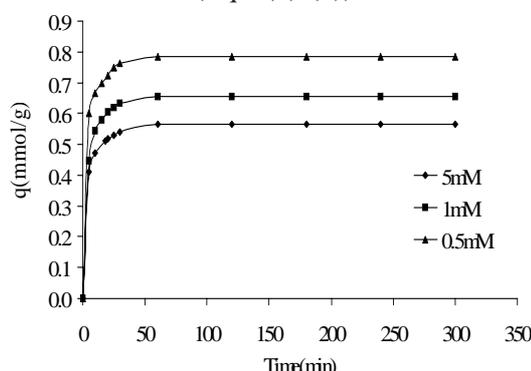


(a)

Figs. 2 (a)-(d) and 3 (a)-(d) show kinetic modeling of Pb<sup>2+</sup> and Cd<sup>2+</sup> biosorption by linear plots of the pseudo first-order, pseudo second-order, saturation and second-order rate equations (Eqs. (1)-(4)). Kinetic parameters of these equations for biosorption of Pb<sup>2+</sup> and Cd<sup>2+</sup> by *Sargassum spp.* biomass were shown in Tables 1 and 2.

### Isotherm study

Fig. 4(a)-(d) shows isotherm modeling of Pb<sup>2+</sup> and Cd<sup>2+</sup> biosorption by linear plots of the Freundlich, Langmuir, Freundlich-Langmuir and Redlich-Peterson models (Eqs. (5)-(8)).



(b)

Fig. 1. Kinetic profiles of (a) Pb<sup>2+</sup> and (b) Cd<sup>2+</sup> biosorption by *Sargassum spp.* biomass

Table 1. Kinetic parameters of the pseudo first-order and pseudo second-order rate equations for the biosorption of Pb<sup>2+</sup> and Cd<sup>2+</sup> by *Sargassum spp.* biomass

Metal ion	C <sub>0</sub> (mM)	Pseudo first-order model			Pseudo second-order model		
		q <sub>e</sub>	k <sub>1</sub>	R <sup>2</sup>	q <sub>e</sub>	k <sub>2</sub>	R <sup>2</sup>
Pb <sup>2+</sup>	5	0.98	0.21	0.879	0.99	1.86	1.000
	1	1.01	0.14	0.839	1.01	0.83	1.000
	0.5	1.07	0.12	0.880	1.08	0.52	1.000
Cd <sup>2+</sup>	5	0.57	0.11	0.377	0.57	1.07	1.000
	1	0.65	0.13	0.817	0.66	0.93	1.000
	0.5	0.79	0.12	0.464	0.79	0.87	1.000

C<sub>0</sub> = initial concentration of metal ion.

R = correlation coefficient.

Table 2. Kinetic parameters of the saturation and second-order rate equations for the biosorption of Pb<sup>2+</sup> and Cd<sup>2+</sup> by *Sargassum spp.* biomass

Metal ion	C <sub>0</sub> (mM)	Saturation model			Second-order model		
		K	k <sub>0</sub>	R <sup>2</sup>	q <sub>e</sub>	k	R <sup>2</sup>
Pb <sup>2+</sup>	5	-3.88	0.0034	0.999	0.98	4.53	0.507
	1	-0.80	0.0010	0.998	1.01	2.43	0.782
	0.5	-0.40	0.0007	0.997	1.07	0.71	0.825
Cd <sup>2+</sup>	5	-4.46	0.0011	1.000	0.57	0.98	0.962
	1	-0.88	0.0004	0.999	0.65	1.41	0.760
	0.5	-0.42	0.0002	1.000	0.79	0.91	0.860

C<sub>0</sub> = initial concentration of metal ion.

R = correlation coefficient.

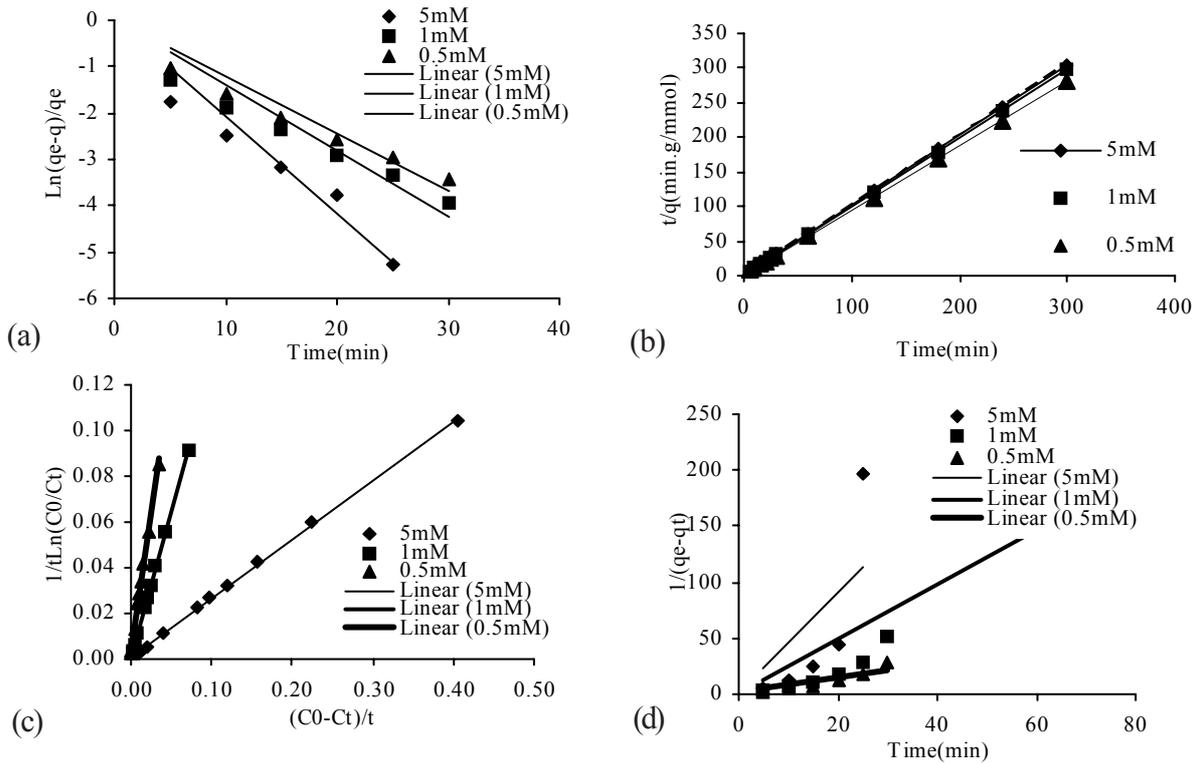


Fig. 2: Kinetic analysis of  $Pb^{2+}$  biosorption by linear plots of (a) the pseudo first-order, (b) pseudo second-order, (c) saturation and (d) second-order rate equations

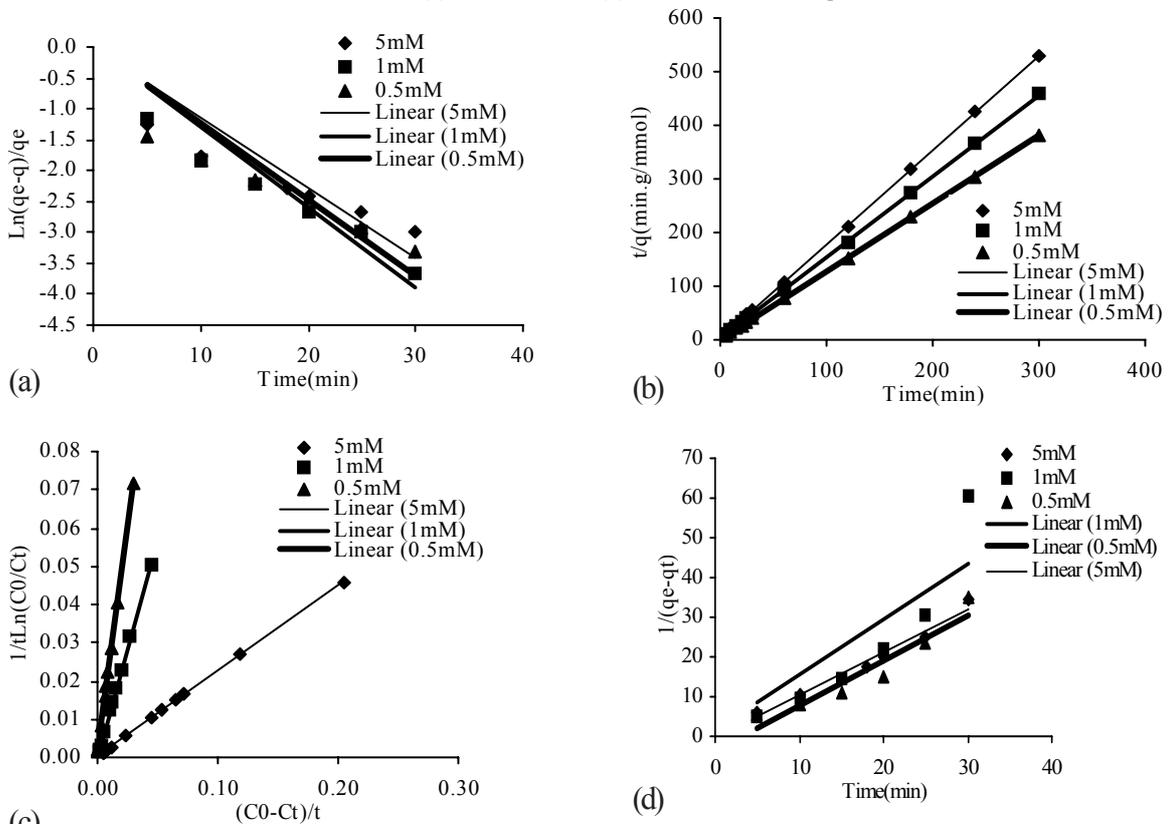


Fig. 3: Kinetic analysis of  $Cd^{2+}$  biosorption by linear plots of (a) the pseudo first-order, (b) pseudo second-order, (c) saturation and (d) second-order rate equations

The maximum biosorption capacities ( $q_m$ ) obtained from this research with those of other biosorbents reported in the literature were given in Table 5.

*Effect of pH on biosorption*

Fig. 5 shows the effect of pH on equilibrium uptake capacities of *Sargassum spp.* biomass for  $Pb^{2+}$  and  $Cd^{2+}$ .

*Effect of light metal ions on biosorption*

The effect of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  on equilibrium capacities of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption by *Sargassum spp.* biomass was shown in Fig. 6.

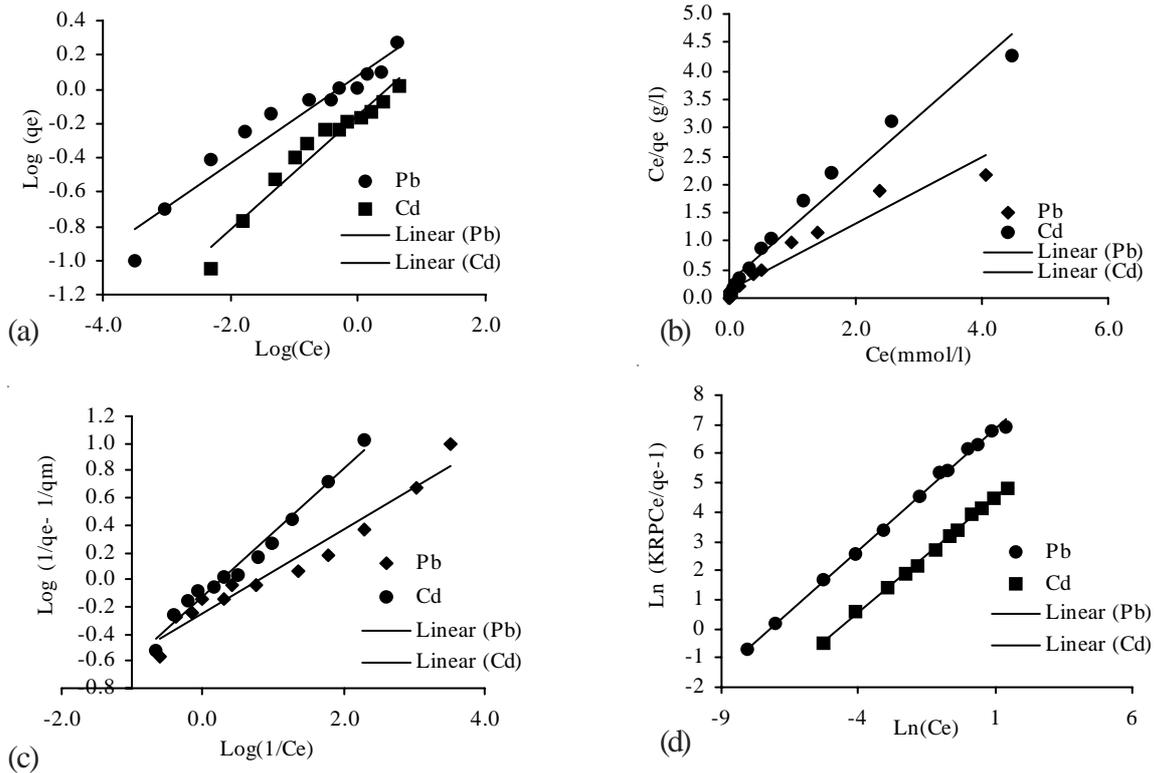


Fig. 4: Isotherm analysis of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption by linear plots of (a) the Freundlich, (b) Langmuir, (c) Freundlich-Langmuir and (d) Redlich-Peterson models

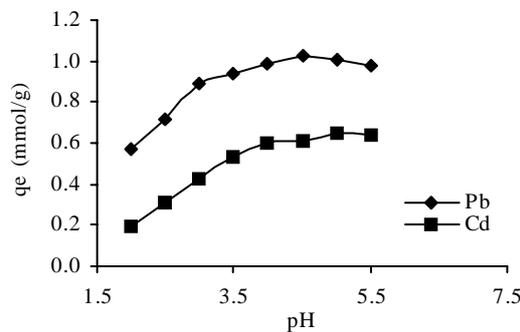


Fig. 5: Effect of pH on equilibrium capacities of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption by *Sargassum spp.* biomass.

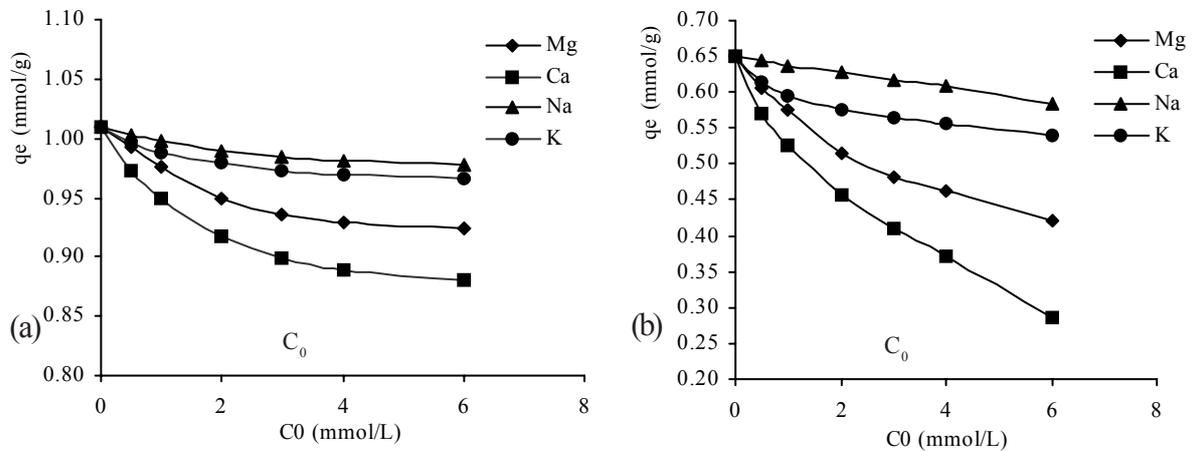


Fig. 6. Effect of  $\text{Na}^+$ ,  $\text{K}^+$ ,  $\text{Mg}^{2+}$  and  $\text{Ca}^{2+}$  on equilibrium capacities of (a)  $\text{Pb}^{2+}$  and (b)  $\text{Cd}^{2+}$  biosorption by *Sargassum spp.* biomass ( $C_0$  = initial concentration of light metal ions).

## DISCUSSION

**Kinetic study:** According to Fig. 1,  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  uptake was relatively fast for all the concentrations studied. At the initial  $\text{Pb}^{2+}$  concentration of 5mM, the system reached to equilibrium within 30min. In general, the heavy metals uptake reached to 88-96% equilibrium capacity of biosorption in 15 min. This rapid kinetic has significant practical importance as it will facilitate smaller reactor volumes ensuring efficiency and economy. Similar rapid metal uptake has been reported for the biosorption of  $\text{Pb}^{2+}$  using *Ecklonia radiata* wherein the system reached over 50-60% of the equilibrium uptake capacity in 10 min (Matheickal and Yu, 1996). The kinetic of chromium (III) biosorption by *Sargassum spp.* biomass was fast, reaching 60% of the total uptake capacity in 10min (Cossich *et al.*, 2002). The biosorption of lead (II) by *Durvillaea potatorum* was rather rapid and 90% of the total uptake occurred in 30min (Matheickal and Yu, 1999). The pseudo second-order and saturation rate equations described the biosorption kinetic of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  with high correlation coefficient ( $R^2 > 0.99$ ) and better than the other equations. Kinetic analysis of  $\text{Pb}^{2+}$ ,  $\text{Cd}^{2+}$ ,  $\text{Ni}^{2+}$  and  $\text{Zn}^{2+}$  biosorption by *Mucor rouxii* represented that the pseudo second-order rate equation described the biosorption kinetic better

than the Langergren model (Yan and Viraraghavan, 2003).

The rate constants of the pseudo second-order rate equation for  $\text{Pb}^{2+}$  biosorption were obtained 1.86, 0.83 and 0.52gm/mol/min at initial  $\text{Pb}^{2+}$  concentrations of 5, 1 and 0.5mM, respectively. In addition, the rate constants of saturation rate equation for  $\text{Pb}^{2+}$  biosorption were determined to be 0.0034, 0.0010 and 0.0007mM/min at initial  $\text{Pb}^{2+}$  concentrations of 5, 1 and 0.5mM, respectively. The rate constants obtained from the pseudo second-order rate equation for  $\text{Cd}^{2+}$  biosorption were 1.07, 0.93 and 0.87gm/mol/min at initial  $\text{Cd}^{2+}$  concentrations of 5, 1 and 0.5 mM, respectively. Also the rate constants of saturation rate equation for  $\text{Cd}^{2+}$  biosorption were determined to be 0.0011, 0.0004 and 0.0002mM/min at initial  $\text{Cd}^{2+}$  concentrations of 5, 1 and 0.5mM, respectively. An increase in initial concentration of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  led to an increase in the rate constant value, therefore there was a direct relationship between initial concentration of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  and the rate of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  biosorption by *Sargassum spp.* biomass. In other words, the biosorption of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  by *Sargassum spp.* was faster in higher initial metal ion concentration.

**Isotherm study:** Isotherm data are basic requirements for the design of biosorption

reactors, moreover analysis of biosorption isotherm is important to develop an equation which accurately represents the results and which can be used for design purposes (Volesky, 2001; Aksu, 2002).

The Redlich-Peterson equation described the isotherm of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption by *Sargassum spp.* biomass with high correlation coefficient ( $R^2 > 0.99$ ) and better than the other models. Also the other models were found in relatively good fitness with the experimental data ( $R^2 > 0.93$ ). According to Langmuir equation, the maximum capacities of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption ( $q_m$ ) were obtained 1.70 and 1.02 mmol/g, respectively. The Langmuir parameter  $q_m$  (maximum uptake capacity) is a suitable measure for comparing different sorbents for the same sorbate. Although due to the various experimental conditions employed in different studies, comparison of their results is difficult, but maximum uptake capacity of *Sargassum spp.* biomass for  $Pb^{2+}$  and  $Cd^{2+}$  far exceed those of most of the biosorbents (Table 5); consequently, the *Sargassum spp.* used in this study can be classified as a good biosorbent.

#### Effect of pH on biosorption

Other studies on heavy metal biosorption have presented that pH was an important parameter affecting the biosorption process (Yan and Viraraghavan, 2003; Selatnia *et al.*, 2004a).

The effect of pH on  $Pb^{2+}$  and  $Cd^{2+}$  biosorption was studied in the initial pH range of 2 to 5.5. At higher pH values, the experiments were not conducted to avoid formation of insoluble  $Pb^{2+}$  and  $Cd^{2+}$  hydroxides. The optimum initial pH values for  $Pb^{2+}$  and  $Cd^{2+}$  biosorption were determined as 4.5 and 5, respectively. The sharpest increase in  $Pb^{2+}$  uptake was obtained between pH 2 and 3. The sharpest increase in  $Cd^{2+}$  uptake was determined between pH values of 2 and 4. The dependence of  $Pb^{2+}$  and  $Cd^{2+}$  biosorption on pH could be largely related to ionic state of binding sites on the algal cell wall (Matheickal and Yu, 1996; Sheng *et al.*, 2004). Measurement of final pH represented the simultaneous release of  $H^+$  with the uptake of heavy metal ions, because final pH of solutions was less than initial pH of solutions, therefore ion exchange confirmed to be one of the biosorption

mechanisms. Other studies with seaweed and fungal biomass have indicated ion exchange as the dominant mechanism of biosorption (Fourest and Roux, 1992; Schiewer and Volesky, 1996; Ahuja *et al.*, 1999; Volesky, 2001).

#### Effect of light metal ions on biosorption

Industrial effluents contaminated with heavy metals contain various kinds of impurities such as light metal ions ( $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$ ) that affect the heavy metal removal processes. Matheickal and Yu (1999) investigated the effect of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  on biosorption of  $Pb^{2+}$  by *Durvillaea potatorum* and *Ecklonia radiata*. The results showed that the biosorbents had much higher relative affinities for  $Pb^{2+}$  than for the light metal ions. The presence of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  in solution did not affect the biosorption capacity of  $Cu^{2+}$  by *Padina spp.* significantly (Kaewsarn, 2002).

The effect of  $Na^+$  and  $K^+$  on  $Pb^{2+}$  uptake by *Sargassum spp.*

biomass was insignificant even at 6 mM concentration of these ions, but  $Mg^{2+}$  and  $Ca^{2+}$  had influence on  $Pb^{2+}$  biosorption. The equilibrium capacity of  $Pb^{2+}$  biosorption was reduced at initial  $Mg^{2+}$  and  $Ca^{2+}$  concentration of 6 mM by 8% and 13%, respectively. The presence of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  in solution affected the biosorption of  $Cd^{2+}$  considerably, so that equilibrium uptake capacity of  $Cd^{2+}$  was reduced at initial concentration 0.5-6 mM of  $Na^+$ ,  $K^+$ ,  $Mg^{2+}$  and  $Ca^{2+}$  by 1-10%, 6-17%, 7-35% and 12-56%, respectively.

Altogether the *Sargassum spp.* used in this study can be classified as an efficient biosorbent because of rapid kinetic, remarkable biosorption capacity and selective removal of metals. Thus the biosorbent has a high potential for application in full-scale for removal of heavy metals from aqueous environments.

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